

**MINERALOGIA**, 42, No 2-3: 75-91 (2011)

DOI: 10.2478/v10002-011-0008-5

[www.Mineralogia.pl](http://www.Mineralogia.pl)

MINERALOGICAL SOCIETY OF POLAND

POLSKIE TOWARZYSTWO MINERALOGICZNE



Original paper

## Pyromorphite formation from montmorillonite adsorbed lead

Tomasz BAJDA<sup>1\*</sup>, Tomasz MARCHLEWSKI<sup>2</sup>, Maciej MANECKI<sup>1</sup>

<sup>1</sup> AGH University of Science and Technology, Faculty of Geology, Geophysics and Environment Protection, al. Mickiewicza 30, 30-059 Krakow, Poland, e-mail: [bajda@geol.agh.edu.pl](mailto:bajda@geol.agh.edu.pl)

<sup>2</sup> Department of Geological Sciences, Miami University, Oxford, Ohio, USA, e-mail: [marchlta@muohio.edu](mailto:marchlta@muohio.edu)

\* Corresponding author

Received: May 25, 2011

Received in revised form: November 27, 2011

Accepted: January 2, 2012

Available online: March 30, 2012

**Abstract.** The reaction of Pb-adsorbed montmorillonite with aqueous solutions of PO<sub>4</sub> and Cl ions results in the decrease in phosphate concentration associated with the formation of a new phase – pyromorphite Pb<sub>5</sub>(PO<sub>4</sub>)<sub>3</sub>Cl. Pyromorphite crystals range in size from hundreds of nm to several tens of μm, depending on the PO<sub>4</sub>, K, and Ca concentrations in the reacting system. A strong ion-exchange effect of K<sup>+</sup> and Ca<sup>2+</sup> cations on desorption of Pb<sup>2+</sup> from Pb-adsorbed montmorillonite was observed. Also, a high concentration of cations leads to a rapid desorption of Pb and the formation of fine pyromorphite crystals. In contrast, low PO<sub>4</sub>, K and Ca concentrations result in the formation of relatively large euhedral crystals. Final Pb concentrations are much lower in experimental sets than in control experiments with no phosphate present.

*Key-words:* sorption, desorption, immobilization, crystallization, apatite, smectite

### 1. Introduction

The addition of phosphate amendments to Pb-contaminated media has been investigated over the past few decades as a method of in-situ lead immobilization. This process aims to change Pb speciation from bioavailable (e.g. aqueous or sorbed) to unavailable for circulation (formation of Pb-phosphates with low solubilities). In this context, the formation of pyromorphite Pb<sub>5</sub>(PO<sub>4</sub>)<sub>3</sub>Cl is especially desirable because of its ease of formation under earth-surface conditions and its low solubility (log K<sub>sp</sub> of -79.66;

Flis et al. 2010). Also, the formation of  $Pb_3(PO_4)_2Cl$  is rapid, reaching steady-state conditions within minutes of the reaction starting. The addition of phosphates to contaminated media has been shown to be an effective method of Pb immobilization and reduction of its bioavailability (Ma et al. 1994; Cotter-Howells 1996; Hettiarachchi et al. 2000, 2001; Ryan et al. 2001; Basta, McGowen 2004; Raicevic et al. 2005; Manecki et al. 2006; Miretzky, Fernandez-Cirelli 2008). It is an effective means of immobilization of Pb in aqueous solutions as well as in contaminated soils (Ma et al. 1993, 1994, 1997). There are many easily available and easy-to-use phosphate sources which can be used for Pb sequestration (e.g., rock-phosphate, fertilizers, apatites; Melamed et al. 2002). Doubts have arisen as to the effectiveness of the application of  $PO_4$  compounds in natural soil systems and as to the stability and bioavailability of Pb in pyromorphite due to microbial activity (Scheckel, Ryan 2002; Chappell, Scheckel 2007; Xie, Giammar 2007; Manecki, Maurice 2008; Debela et al. 2010). Owing to their diversity, the application of different  $PO_4$  compounds introduces differing amounts of cations associated with  $PO_4$  as well as differing proportions of the cations and  $PO_4$  to Pb. Thus, an understanding the influence of these variables on the mechanisms and effectiveness of Pb sequestration would allow for the more efficient treatment of contaminated sites.

The solid speciation of Pb in soils can vary greatly depending on the soil type and its anthropogenic history. Most soil lead is associated with Pb-minerals and sorbed on organic matter, clay minerals and Fe, Mn, Al oxides (Sauvé et al. 2000). Therefore, studies of Pb immobilization must take into account its speciation. Most previous research has focused on the reactions between apatites and aqueous Pb (Maneckí et al. 2000 and literature therein), i.e. between Pb minerals (anglesite, cerussite, galena) and aqueous phosphates or apatites (Ruby et al. 1994; Zhang, Ryan 1998, 1999a, 1999b; Stack et al. 2004) and between phosphate-treated goethite and kaolinite and aqueous Pb (Maneckí et al. 2006; Taylor et al. 2009). Zhang et al. (1997) suggested that it is possible, by pyromorphite formation, to stabilize Pb adsorbed on goethite. However, little is yet known about the mechanisms of sorbed-Pb immobilization by phosphate treatments.

This study focuses on the mechanisms of pyromorphite formation during the reactions between Pb-adsorbed montmorillonite and aqueous solutions with different concentrations of  $PO_4$ , K, and Ca. Montmorillonite is one of the most widespread phyllosilicate minerals in soils. It also has large cation adsorption capacity. Hence, it may play an important role in Pb transport and immobilization processes (Ranatunga et al. 2008). Montmorillonite surface- and Pb adsorption mechanism differ from those of Al and Fe oxides. Thus, the behavior and fate of Pb adsorbed on these phases may be different. This paper illustrates the mechanisms of pyromorphite formation during immobilization reactions between  $PO_4$  and Pb-adsorbed montmorillonite.

## **2. Materials and methods**

### **2.1. Materials**

The source of montmorillonite used in the experiments was bentonite from Jelšový Potok, Slovakia. The clay-size fraction ( $< 2 \mu m$ ) was separated from bentonite by a sedimentation procedure. The mineralogical composition was determined by X-ray

powder diffraction (XRD) in conjunction with standard cation-saturation procedures for the identification of expandable phyllosilicates (Moore, Reynolds 1997). The separated smectite concentrate consisted only of montmorillonite (Fig. 1). The XRD analysis showed no evidence of any other phases, confirming that the sedimentation process had effectively removed all mineral impurities from the initial sediment. The XRD analysis of the air-dried sample gave a very strong 001 diffraction peak at 14.9 Å which, in the glycol-treated sample, shifts to 17.0 Å. This indicates that  $\text{Ca}^{2+}$  is the predominant cation in the smectite interlayer positions (Moore, Reynolds 1997). Final identification was accomplished by drying the sample at 300°C. This treatment shifted the 001 diffraction peak to 9.6 Å (Fig. 1), indicative of montmorillonite collapse. Results of the analysis of the initial montmorillonite composition obtained by wavelength dispersive X-ray fluorescence spectroscopy (Philips PW, 1410, Philips, Holland; Si, Al, Fe, Mg, Ca) and atomic absorption spectrometry (AAS; Na, K, Ti) are as follows (in wt%):  $\text{SiO}_2$  – 67.25,  $\text{Al}_2\text{O}_3$  – 21.85,  $\text{Fe}_2\text{O}_3$  – 2.82,  $\text{MgO}$  – 3.64,  $\text{CaO}$  – 0.15,  $\text{Na}_2\text{O}$  – 0.24,  $\text{K}_2\text{O}$  – 0.15,  $\text{TiO}_2$  – 0.12 and  $\text{H}_2\text{O}$  – 4.08. The cation exchange capacity (CEC) determined by sorption of hexadecyltrimethylammonium (HDTMA) bromide is 1.04 meq/g. To determine the amount of HDTMA bound by the mineral, C and N contents in the organo-smectite were measured

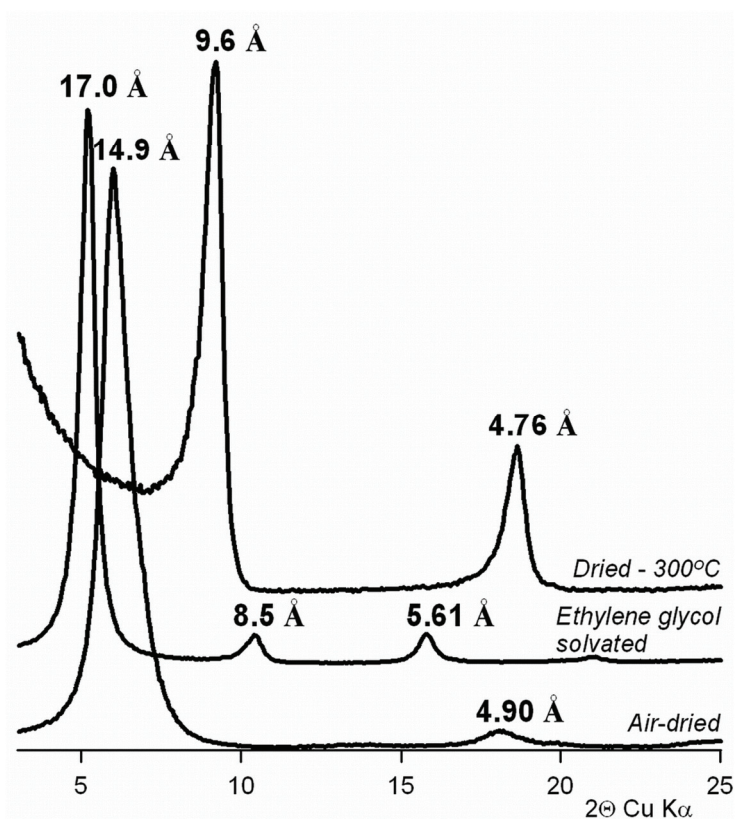


Fig. 1. Montmorillonite used in the study in the air-dried-, 300°C dried- and ethylene glycol-solvated states.

using a Euro EA elemental analyzer (HEKAtech GmbH, Germany). The BET surface area measured on an ASAP 2405 automatic sorption analyzer from Micromeritics Inc. (Micromeritics, USA), using the N<sub>2</sub> adsorption method, was 715 m<sup>2</sup>/g. The synthetic pyromorphite used as a reference material in the XRD study is described in our previous work (Flis et al. 2010).

## 2.2. Pb sorption

To prepare Pb-adsorbed montmorillonite, a suspension of 50 g/L montmorillonite was reacted with a 1 M Pb(NO<sub>3</sub>)<sub>2</sub> solution in the presence of 0.01 M Ca(NO<sub>3</sub>)<sub>2</sub> background electrolyte at an initial pH of 3.5. The pH of the solution was not adjusted and not maintained during the sorption process. The suspension was stirred for 24 hours using a mechanical stirrer and then left for a further 24 hours to reach equilibrium. Based on studies of the influence of time on the amount of metals adsorbed on montmorillonite, it was found that 24 h is a sufficient period for attaining the equilibrium state between Pb ions in the solution and adsorbed on smectite (Mozgawa et al. 2009). After equilibrium was reached, the Pb-adsorbed montmorillonite suspension was centrifuged and placed in dialysis tubes to remove all the ions remaining in the intergranular solution (Pb, Ca, NO<sub>3</sub>). The tubes had a molecular weight cutoff of 1000 Da, allowing free passage of inorganic ions. The dialysis tubes were then placed in redistilled water, which was replaced every 24 hours, and its Pb concentration was measured. After 60 days, when the Pb concentration in the solution was less than 0.1 ppm, the Pb-montmorillonite suspension was freeze dried using an Alpha 1-2 LD (CHRIST, Germany) drier.

The amount of exchangeable lead was determined by a desorption experiment. 1 g of freeze dried Pb-adsorbed montmorillonite was mixed with 5 mL of 1 M CH<sub>3</sub>COONH<sub>4</sub> solution (pH = 7). The suspension was shaken for 2 hours, then centrifuged and 4 mL of supernatant solution was sampled and replaced by the same volume of fresh 1 M CH<sub>3</sub>COONH<sub>4</sub> solution. This process was repeated three times. The decanted solution was centrifuged, filtered using a 0.1 µm polycarbonate filter and the Pb concentration was measured by atomic absorption spectrometry (AAS). The determined amount of sorbed Pb was 0.35 ± 0.03 mmol Pb per g of montmorillonite (0.7 ± 0.06 meq Pb/g). The fact that the CEC of montmorillonite was equal to 1.04 meq/g indicates that the ion exchange sites of the montmorillonite were not fully exchanged with Pb ions.

## 2.3. Experiments

The purpose of the experiments was to investigate the mechanisms of the reaction between Pb-adsorbed montmorillonite and phosphate amendments during remediation by in-situ immobilization. In all experiments, 500 mg of montmorillonite was added to polypropylene bottles containing 500 ml of phosphate-additive solution. The pH of the suspensions was adjusted with 0.1 M HNO<sub>3</sub> or 0.1 M KOH to a pH of 5.0 ± 0.1. During the reactions, the pH was maintained and adjusted every day for two weeks. At the beginning of each reaction, the suspension was intensively shaken for 5 min and then stirred every day. 10 ml solution aliquots were taken for analysis after 1, 3, 7 and 14 days of reaction, centrifuged and filtered using a 0.1 µm polycarbonate filter. Concentrations of K, Pb

(AAS), Ca (ICP-OES), PO<sub>4</sub> (UV-Vis spectrophotometry) and pH were measured. At the end of each experiment, the suspensions were filtered through a 0.45 μm membrane, and the collected solids washed with redistilled water and acetone before air-drying for XRD and SEM-EDS analysis.

Two sets of experiments were conducted. In set 1, Pb-adsorbed montmorillonite was reacted with aqueous solutions containing K, PO<sub>4</sub> and Cl to simulate fertilizer phosphate amendments to a Pb-contaminated soil. In set 2, Pb-adsorbed montmorillonite was reacted with solutions containing Ca, K, PO<sub>4</sub> and Cl to simulate the apatite amendments commonly used in remediation procedures. The preliminary assumption was that all adsorbed Pb will be desorbed during the reactions and incorporated in precipitating pyromorphite Pb<sub>5</sub>(PO<sub>4</sub>)<sub>3</sub>Cl. The stoichiometry of this reaction determined the concentrations of aqueous solutions used. For example, a 0.21 mM PO<sub>4</sub> (as KH<sub>2</sub>PO<sub>4</sub>) solution was used based on (a) the molar ratio Pb : PO<sub>4</sub> = 1.67 in pyromorphite, (b) the mass of Pb-adsorbed montmorillonite used in the experiments (500 mg in 500 ml of solution) and (c) the content of Pb adsorbed on the montmorillonite (0.35 mmol Pb/g). Each experimental set contained three series with different K : Pb : PO<sub>4</sub> (set 1) or K : Ca : Pb : PO<sub>4</sub> (set 2) proportions (Tables 1 and 2):

Series 1: 0.1 M K (KCl) or 0.1 M Ca (CaCl<sub>2</sub>) concentrations and 0.21 mM PO<sub>4</sub> (KH<sub>2</sub>PO<sub>4</sub>) according to the stoichiometry of pyromorphite (PO<sub>4</sub> : Pb = 3 : 5).

Series 2: ten times greater concentration of PO<sub>4</sub> (KH<sub>2</sub>PO<sub>4</sub>) than in series 1 (PO<sub>4</sub> : Pb = 6 : 1); Ca (CaCl<sub>2</sub>) concentration equal to the amount of adsorbed Pb; 2.11 mM K to reflect its content in the phosphate source (KH<sub>2</sub>PO<sub>4</sub>); 0.07 mM Cl (KCl) according to the stoichiometry of pyromorphite relative to the adsorbed Pb (Cl : Pb = 1 : 5).

Series 3: the amount of cations (K, Ca) equal to the amount of adsorbed Pb (0.35 mM g<sup>-1</sup>); 0.21 mM PO<sub>4</sub> (KH<sub>2</sub>PO<sub>4</sub>) in stoichiometric proportions relative to the adsorbed Pb (PO<sub>4</sub> : Pb = 3 : 5).

TABLE 1

Assumed concentrations of K, Cl, and PO<sub>4</sub> in set 1.

	Experiment			Control		
Series 1 [mM]	K	PO <sub>4</sub>	Cl	K	PO <sub>4</sub>	Cl
	100.21	0.21	100.00	100.21	0	100.00
Molar ratio	Pb : K	Pb : PO <sub>4</sub>	Pb : Cl	Pb : K	-	Pb : Cl
	1 : 286	5 : 3	1 : 286	1 : 286	-	1 : 286
Series 2 [mM]	K	PO <sub>4</sub>	Cl	K	PO <sub>4</sub>	Cl
	2.18	2.11	0.07	2.18	0	0.07
Molar ratio	Pb : K	Pb : PO <sub>4</sub>	Pb : Cl	Pb : K	-	Pb : Cl
	1 : 6	1 : 6	5 : 1	1 : 6	-	5 : 1
Series 3 [mM]	K	PO <sub>4</sub>	Cl	K	PO <sub>4</sub>	Cl
	0.35	0.21	0.14	0.35	0	0.14
Molar ratio	Pb : K	Pb : PO <sub>4</sub>	Pb : Cl	Pb : K	-	Pb : Cl
	1 : 1	5 : 3	2.5 : 1	1 : 1	-	2.5 : 1

TABLE 2

Assumed concentrations of Ca, K, Cl, and PO<sub>4</sub> in set 2.

	Experiment				Control			
Series 1	Ca	K	PO <sub>4</sub>	Cl	Ca	K	PO <sub>4</sub>	Cl
[mM]	100.00	0.21	0.21	200.00	100.00	0.21	0	200.00
Molar ratio	Pb : Ca	Pb : K	Pb : PO <sub>4</sub>	Pb : Cl	Pb : Ca	Pb : K	-	Pb : Cl
	1 : 284	5 : 3	5 : 3	1 : 512	1 : 284	5 : 3	-	1 : 512
Series 2	Ca	K	PO <sub>4</sub>	Cl	Ca	K	PO <sub>4</sub>	Cl
[mM]	0.35	2.11	2.11	0.70	0.35	2.11	0	0.70
Molar ratio	Pb : Ca	Pb : K	Pb : PO <sub>4</sub>	Pb : Cl	Pb : Ca	Pb : K	-	Pb : Cl
	1 : 1	1 : 6	1 : 6	1 : 2	1 : 1	1 : 6	-	1 : 2
Series 3	Ca	K	PO <sub>4</sub>	Cl	Ca	K	PO <sub>4</sub>	Cl
[mM]	0.35	0.21	0.21	0.70	0.35	0.21	0	0.70
Molar ratio	Pb : Ca	Pb : K	Pb : PO <sub>4</sub>	Pb : Cl	Pb : Ca	Pb : K	-	Pb : Cl
	1 : 1	5 : 3	5 : 3	1 : 2	1 : 1	5 : 3	-	1 : 2

Along with the experimental sets, control samples were prepared. In the control samples, aqueous solutions of cations with concentrations consistent with those in the corresponding experimental samples were reacted with Pb-adsorbed montmorillonite in the absence of PO<sub>4</sub>. An additional control experiment was carried out in which Pb-free montmorillonite was reacted with 0.21 mmol/L PO<sub>4</sub> solution at pH 5.0 (Fig. 2).

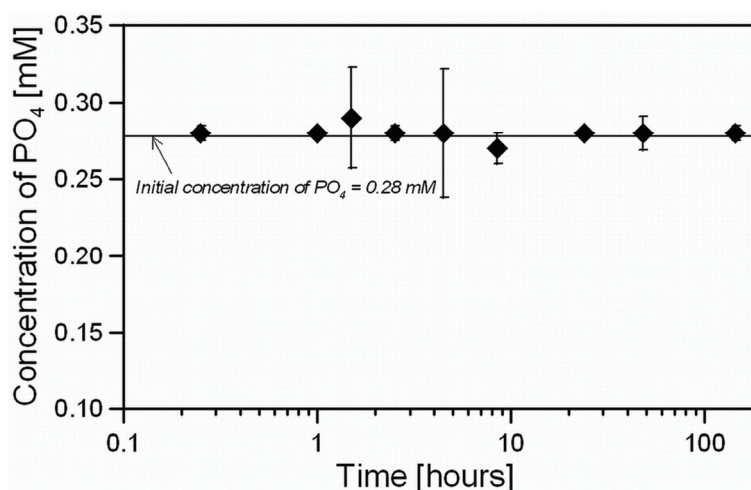


Fig. 2. Amount of phosphate sorbed by Ca-montmorillonite at 0.28 mM PO<sub>4</sub> initial concentration.

The reactions were considered to have reached equilibrium when concentrations of Pb and PO<sub>4</sub> in three consecutive solution samples were equal within two standard deviations of triplicate results (experimental error). The *t*-test was used to verify the absence of any trend

in the data from three consecutive analyses and the slope was not significantly different from zero at the 95% significance level.

## 2.4. Analytical Methods

Concentrations of Pb and K in aqueous solutions were determined using a Philips PU-9100x (Phillips, U.K.) atomic absorption spectrometer equipped with a flame head (FAA) and Ca-using inductively coupled plasma spectrometer (Perkin Elmer Plasm 40; PerkinElmer, USA). PO<sub>4</sub> concentrations were determined colourimetrically by the molybdenum blue method (Lenoble et al. 2003) using a Hitachi 1600 (Hitachi, Japan) spectrophotometer at 870 nm wavelength. The analytical detection limits for Pb, Ca and PO<sub>4</sub> were 5 µg/L and for K – 10 µg/L.

All solids were analyzed using an X-ray powder diffractometer (Philips PW 3020 X'Pert-APD Diffractometer system with graphite monochromator, Philips, Holland) using CuK $\alpha$  radiation, step scan mode with a step size of 0.02° 2 $\theta$  and a rate of 1 s per step. Air-dried uncoated samples were analyzed by electron microscopy using a variable pressure field-emission scanning electron microscope (FE-SEM, Zeiss Supra 35VP, Carl Zeiss, USA) equipped with an energy dispersive spectrometer (EDS) for elemental microanalysis.

## 3. Results and Discussion

The amount of Pb released from Pb-adsorbed montmorillonite to the solution in the control sets (without PO<sub>4</sub>) is positively correlated with the initial concentration of K and Ca. Almost 100% of the sorbed Pb is exchanged by 0.1 M K and 0.17 M Ca+K in series 1 of both experimental sets and up to 50% of the sorbed Pb is desorbed in the other series (Tables 3 and 4). The molar ratios of adsorbed K and Ca to desorbed Pb indicate the ion exchange mechanism (Tables 3 and 4). The amount of Pb desorption in the experiments of set 2 (K and Ca present) is greater than in set 1 (only K present) because of the charge in desorbing cations. To preserve the charge balance in montmorillonite, one divalent Ca can replace one Pb from an ion-exchange site, whereas two monovalent K are necessary for the replacement of one Pb.

In the experimental series 2 of both sets, the presence of phosphates resulted in the reduction of Pb concentrations to about 0.001 mM (Tables 3 and 4). This means that 100% of the desorbed Pb was removed from the solution. In series 1 and 3, the reduction of the aqueous Pb with respect to a control sample was 80–85% and 20–60%, respectively. Pyromorphite Pb<sub>5</sub>(PO<sub>4</sub>)<sub>3</sub>Cl was the only reaction product detected in the experiments (Figs. 3 and 4). A decrease in PO<sub>4</sub> concentration was observed also. This and the lack of sorption of PO<sub>4</sub> on montmorillonite (Fig. 2) confirm that the drop in PO<sub>4</sub> and Pb concentrations is exclusively caused by precipitation of pyromorphite. Therefore, the molar ratio of the decrease in Pb concentration (Pb<sub>removed</sub> – the difference between Pb concentrations in control samples and experimental samples) to the decrease in PO<sub>4</sub> concentration (PO<sub>4removed</sub> – the difference between the initial and the final PO<sub>4</sub> concentrations in the experimental samples) should be consistent with the stoichiometry of pyromorphite Pb<sub>5</sub>(PO<sub>4</sub>)<sub>3</sub>Cl (5 : 3 = 1.67). However, the ratio varied in the series (Table 5). Only in series 1 of both sets are the ratios Pb<sub>removed</sub> : PO<sub>4removed</sub> similar (within experimental error) to the expected 1.67. This

means that the amount of Pb desorbed from montmorillonite (as determined by controls) was sufficient to account for PO<sub>4</sub> loss in the experimental sample. In series 2 and 3, the ratios Pb<sub>removed</sub> : PO<sub>4removed</sub> are < 1.67 indicating that PO<sub>4</sub> loss in the experimental samples was higher than calculated based on Pb desorption in controls and the stoichiometry of pyromorphite. Higher PO<sub>4</sub> loss suggests higher Pb desorption than in the control samples since pyromorphite formation is the only sink for phosphates. This is illustrated in Figures 5 and 6 which show the comparison between the amount of Pb desorbed from montmorillonite in the controls minus the final amount of Pb in the experimental samples and the amount of Pb necessary to form pyromorphite from the amount of PO<sub>4</sub> removed from the solution, (determined by the fall in PO<sub>4</sub> concentration and calculated from pyromorphite stoichiometry). It can be seen that only in series 1 is the amount of Pb desorbed in the control sample sufficient to account for PO<sub>4</sub> loss resulting from the precipitation of pyromorphite. In series 3, the concentration of Pb desorbed from the control equals the Pb concentration in the experimental sample even though pyromorphite formed (Figs 3 and 4).

TABLE 3

Initial and final concentrations in the experiments at various initial K and PO<sub>4</sub> concentrations (set 1)

		Experiment		Control	
		Initial	Equilibrium	Initial	Equilibrium
Series 1 [mM]	K	120.0 ± 0.2	116.1 ± 0.7	117.6 ± 2.2	114.0 ± 2.7
	PO <sub>4</sub>	0.240 ± 0.049	0.022 ± 0.015	–	–
	Pb	–	0.062 ± 0.013	–	0.405 ± 0.010
Ratio K/Pb <sup>A</sup>				3.6 ± 3.5 : 0.40 ± 0.01	
Series 2 [mM]	K	1.767 ± 0.007	1.613 ± 0.002	1.785 ± 0.002	1.736 ± 0.002
	PO <sub>4</sub>	2.508 ± 0.038	2.288 ± 0.056	–	–
	Pb	–	< 0.001	–	0.080 ± 0.002
Ratio K/Pb <sup>A</sup>				0.06 ± 0.00 : 0.08 ± 0.00	
Series 3 [mM]	K	0.301 ± 0.001	0.221 ± 0.007	0.332 ± 0.029	0.276 ± 0.025
	PO <sub>4</sub>	0.206 ± 0.002	0.083 ± 0.002	–	–
	Pb	–	0.024 ± 0.002	–	0.029 ± 0.003
Ratio K/Pb <sup>A</sup>				0.05 ± 0.04 : 0.03 ± 0.00	

The reported concentrations represent the mean ± the two standard deviation from triplicate experiments

<sup>A</sup> Proportion of adsorbed K to desorbed Pb



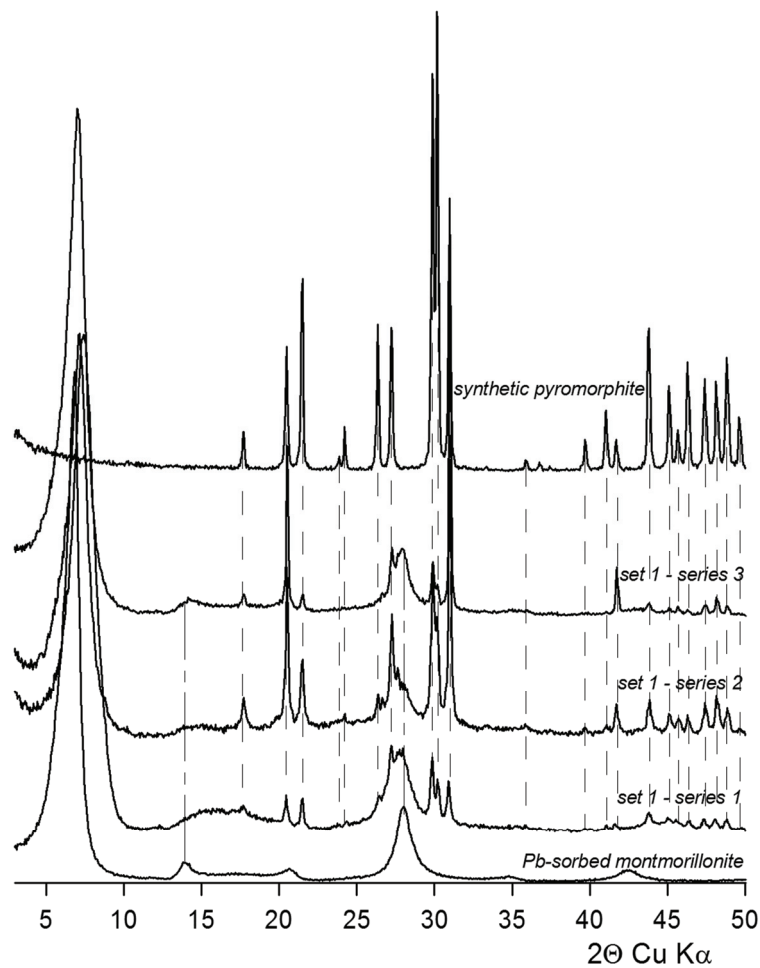


Fig. 3. X-ray diffraction patterns for synthetic pyromorphite and Pb-adsorbed montmorillonite before and after reaction with the solutions containing Pb, PO<sub>4</sub>, K, and Cl.

The discrepancy between the observed amounts of Pb desorption in the control samples and the PO<sub>4</sub> loss in the corresponding experimental samples can be explained by the following reaction mechanism. Cation exchange in the interlayer spaces of montmorillonite is the driving force of Pb desorption from the clay and its appearance in the solution. Therefore, increased aqueous K or Ca concentrations in the reaction system lead to a greater amount of desorbed Pb. Addition of PO<sub>4</sub> in the presence of Cl causes formation of pyromorphite. Due to the very low solubility of pyromorphite, almost all Pb present in the aqueous solutions is incorporated into this mineral. Pb disappears from the solution as a result and a new equilibrium for the reaction is established. If K or Ca is still present in the solution, it exchanges further with Pb from the interlayer spaces of the smectite. Released Pb reacts with PO<sub>4</sub> and a new equilibrium is established again. The approximate amount of Pb needed to form pyromorphite, determined by the drop in PO<sub>4</sub>

concentration (calculated from pyromorphite stoichiometry) in the experimental series 1 and 2 is about 0.4 mM (Figs 5 and 6). This means that all Pb adsorbed on montmorillonite was desorbed and precipitated by  $\text{PO}_4$  to form  $\text{Pb}_5(\text{PO}_4)_3\text{Cl}$ . Only in the experimental series 3 was only approximately 50% adsorbed Pb subsequently desorbed. The reaction between Pb and  $\text{PO}_4$  explains the greater drop in K concentration in the experimental samples than in the controls (Fig. 7), as more Pb is replaced by K in the experimental samples.

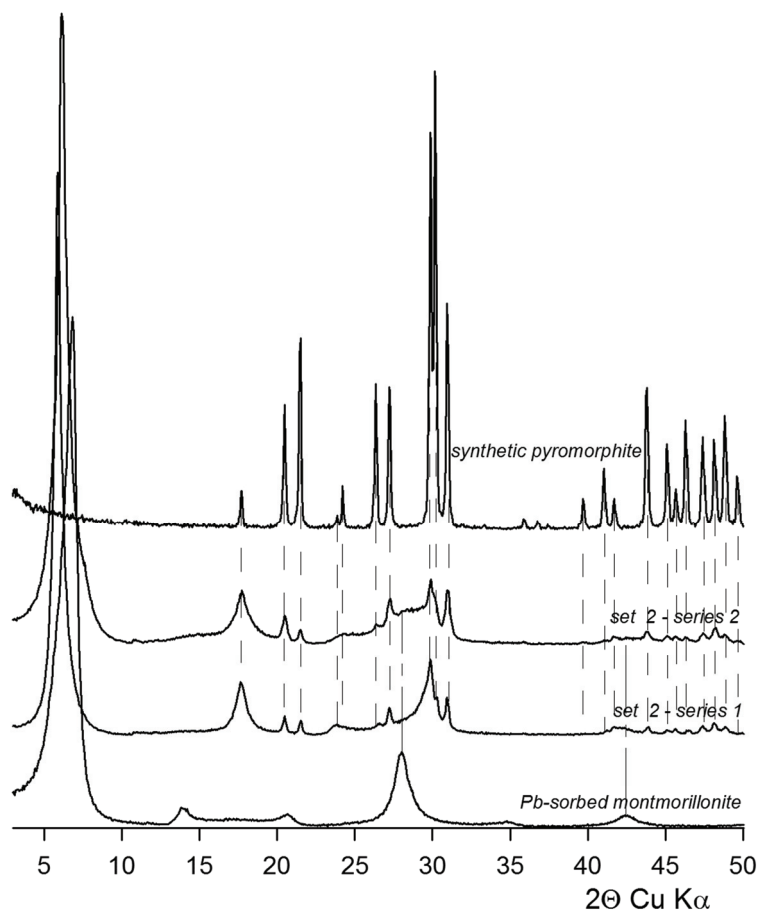


Fig. 4. X-ray diffraction patterns for synthetic pyromorphite and Pb-adsorbed montmorillonite before and after reaction with the solutions containing Pb,  $\text{PO}_4$ , Ca, K, and Cl.

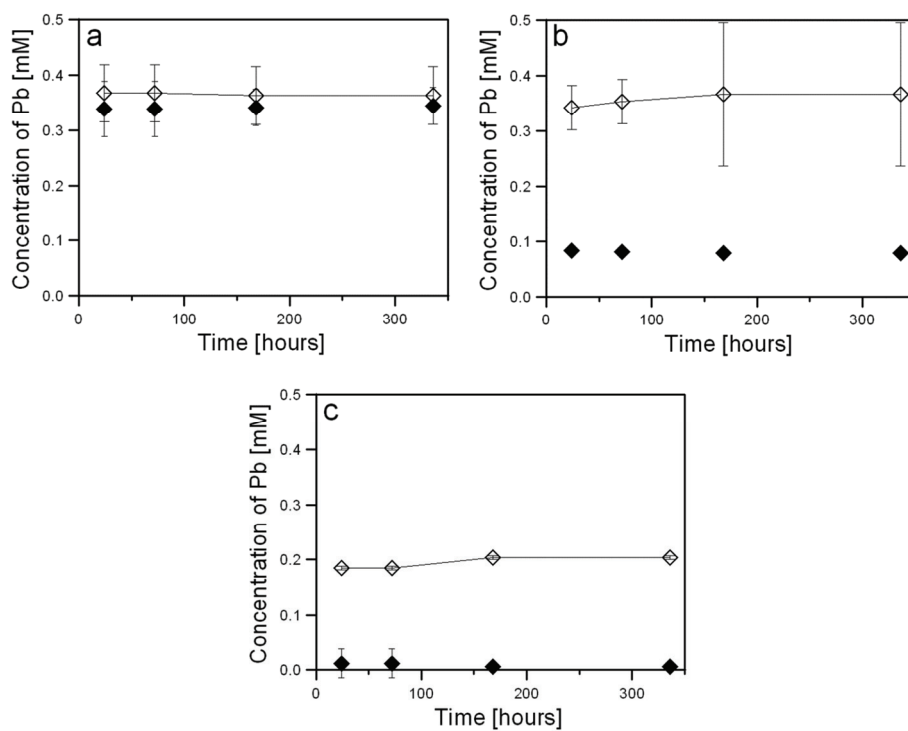


Fig. 5. Comparison of the amount of Pb removed from the solution in set 1 (filled diamonds) to the amount of Pb determined by the drop in PO<sub>4</sub> concentration (calculated from pyromorphite stoichiometry, open diamonds): a) series 1, b) series 2, c) series 3.

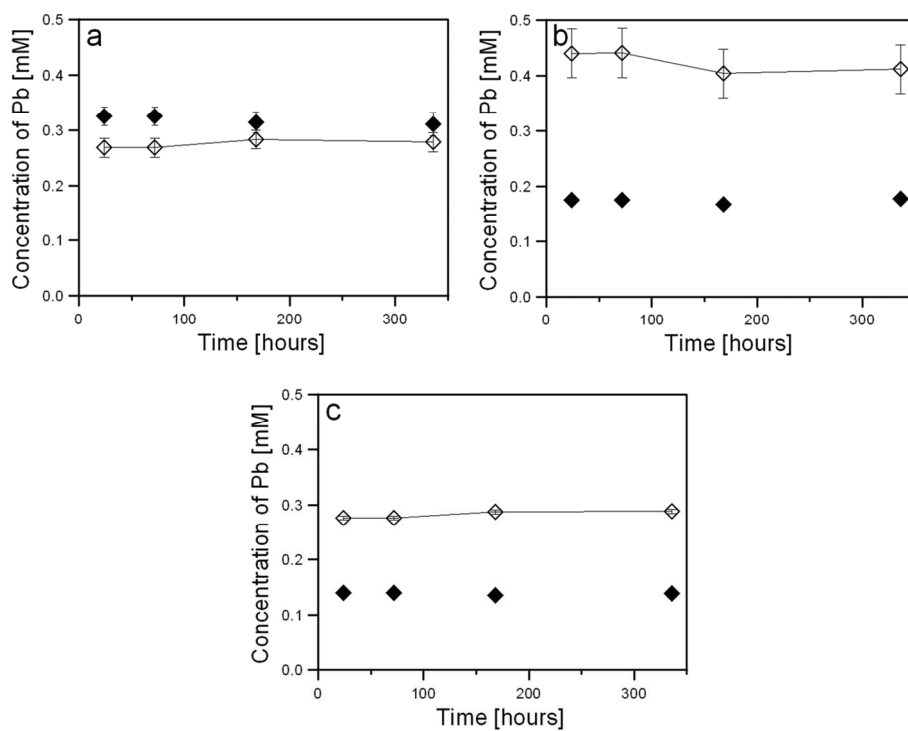


Fig. 6. Comparison of the amount of Pb removed from the solution in set 2 (filled diamonds) to the amount of Pb determined by the drop in PO<sub>4</sub> concentration (calculated from pyromorphite stoichiometry, open diamonds): a) series 1, b) series 2, c) series 3.

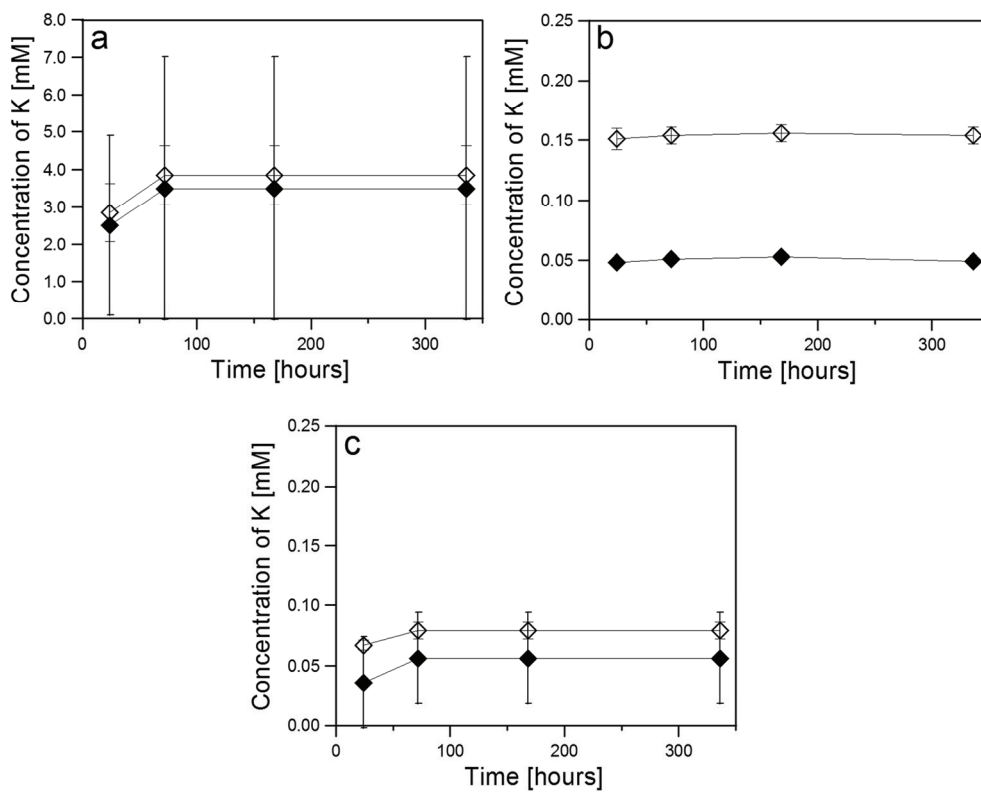


Fig. 7. Fall in concentration of K in experimental samples (open diamonds) and control samples (filled diamonds) for set 1: a) series 1, b) series 2, c) series 3.

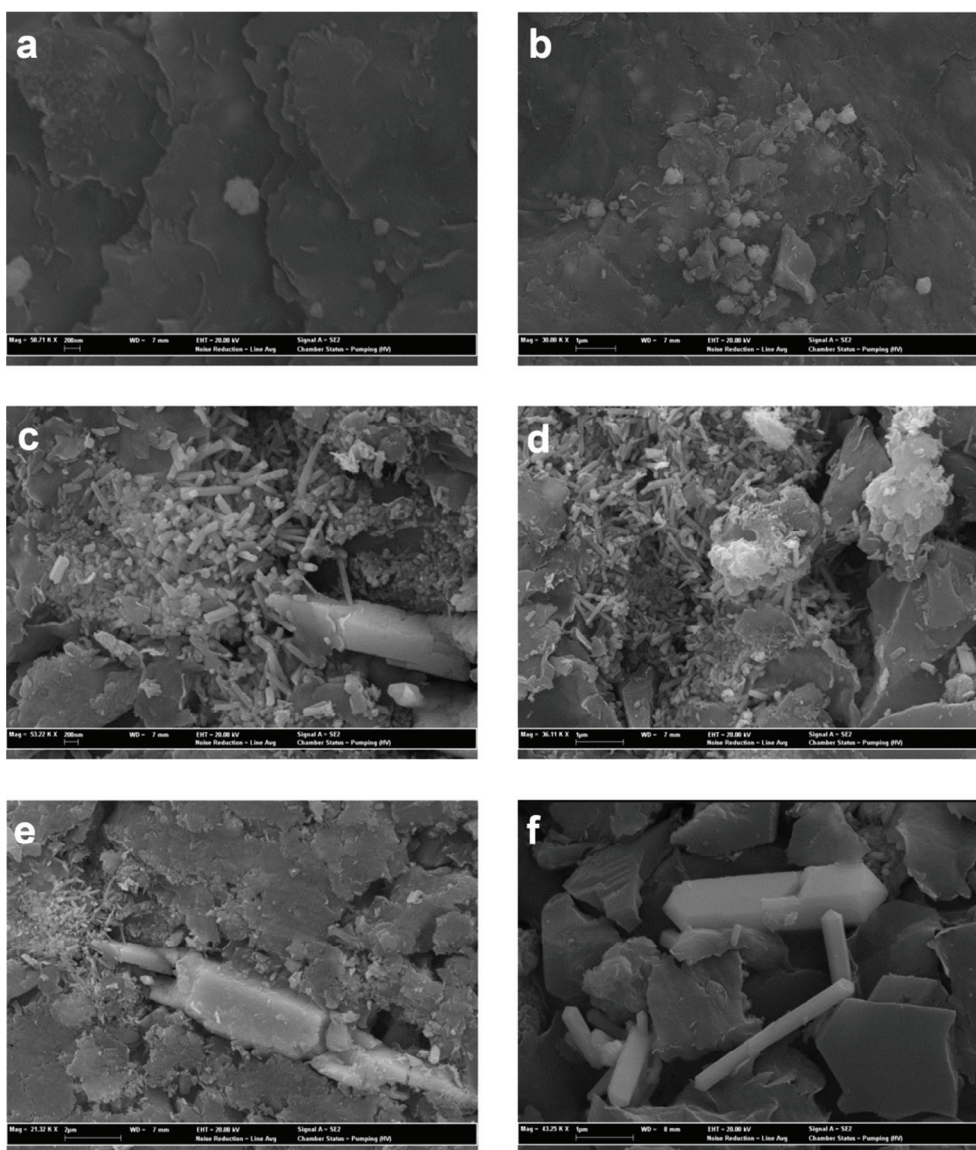


Fig. 8. SEM images of pyromorphite on the surface of montmorillonite: a-b) series 1, c-e) series 2, f) series 3.

The amount of Pb desorption from montmorillonite is controlled by two factors, namely, the aqueous concentration of desorbing cations and the activity of aqueous Pb. Pyromorphite formation shifts the equilibrium between the adsorbed and aqueous Pb, as the latter is being incorporated into the mineral structure. This, in turn, enables more desorption.

The reactions between phosphates and Pb desorbed from montmorillonite are rapid, as equilibrium is reached within 24 hours of the reaction. Cation exchange on

montmorillonite and crystallization of pyromorphite are simultaneous, suggesting that Pb desorption is controlling the precipitation of  $\text{Pb}_5(\text{PO}_4)_3\text{Cl}$ . However, the morphology of pyromorphite crystals and their size varied in each series. Pb-adsorbed montmorillonite reacted with the solutions of series 1 produced very small crystals without well defined faces (Figs 8a and 8b). These tended to form crystal aggregates rather than single crystals. Their morphology does not depend on the type of cation (K, Ca) in the initial solution. The sizes of the aggregates reached up to 500 nm, but were typically smaller. In some cases, they covered the surfaces of smectite plates. In series 2, it is apparent that two generations of pyromorphite crystals formed (Fig. 8c). The first are small crystals that form aggregates similar to the series 1 examples (Fig. 8c), but also some bigger crystals (Fig. 8d). The lengths of the first-generation crystals vary from 300 nm to 1  $\mu\text{m}$ . Second generation crystals show well-defined morphologies with well-formed terminations at both ends that suggests homogenous formation (Fig. 8e). Their sizes vary from several to several tens of  $\mu\text{m}$ . In series 3, only one form of pyromorphite crystals formed (Fig. 8f). These crystals are euhedral and 0.5-3  $\mu\text{m}$  long.

Although the formation of pyromorphite from Pb-adsorbed montmorillonite is quite rapid, the morphological variety of the crystals may suggest different reaction rates. The very fine crystals formed in the series 1 experiments suggest that Pb desorption is very fast with the desorbed Pb being immediately bonded to  $\text{PO}_4$  and Cl to form pyromorphite. Because Pb was exchanged with K or Ca rapidly, a large amount of Pb in the solution led to the formation of many very small crystals. In the series where K and Ca concentrations were lower, Pb was released to the solution less rapidly and equilibrium of the reaction was established more slowly, allowing larger crystals to form.

#### 4. Conclusions

Pb adsorbed on montmorillonite can be easily mobilized via cation exchange by K and Ca ions. The reaction of Pb-adsorbed montmorillonite with aqueous solutions containing  $\text{PO}_4$  ions results in the formation of pyromorphite  $\text{Pb}_5(\text{PO}_4)_3\text{Cl}$ . Pb locked in the pyromorphite structure is no longer a hazard to the environment because of the very low solubility of the mineral ( $\log K_{\text{sp}} = -79.66$ ). The amount of adsorbed Pb that can be immobilized depends on the amount of phosphate available as well as on cations which can exchange with Pb in the interlayer spaces of the clay mineral or on its external surface. Pyromorphite precipitates homogeneously in the solution but crystals generally cover the surface of the smectite. A high concentration of K and Ca cations leads to the formation of very fine crystals. This also causes a greater desorption of Pb. Conversely, low concentrations of cations result in the growth of larger crystals.

*Acknowledgements.* This paper greatly benefited from the analytical help of, and discussions with, Dr. O. Borkiewicz (Argonne National Laboratory) and A. Gawel (AGH UST). The work was supported by Ministry of Science and Higher Education grant no. N N525 461236.

## 5. References

- Basta, N.T., & McGowen, S.L. (2004). Evaluation of chemical immobilization treatments for reducing heavy metal transport in a smelter-contaminated soil. *Environmental Pollution*, 127(1), 73-82. DOI: 10.1016/S0269-7491(03)00250-1.
- Chappell, M.A., & Scheckel, K.G. (2007). Pyromorphite formation and stability after quick lime neutralisation in the presence of soil and clay sorbents. *Environmental Chemistry*, 4(2), 109-113. DOI: 10.1071/EN06081.
- Cotter-Howells, J. (1996). Lead phosphate formation in soils. *Environmental Pollution*, 93(1), 9-16. DOI: 10.1016/0269-7491(96)00020-6.
- Debela, F., Arocena, J.M., Thring, R.W., & Whitcombe, T. (2010). Organic acid-induced release of lead from pyromorphite and its relevance to reclamation of Pb-contaminated soils. *Chemosphere*, 80(4), 450-456. DOI: 10.1016/j.chemosphere.2010.04.025.
- Flis, J., Borkiewicz, O., Bajda, T., Manecki, M., & Klasa, J. (2010). Synchrotron-based X-ray diffraction of the lead apatite series  $Pb_{10}(PO_4)_6Cl_2$ - $Pb_{10}(AsO_4)_6Cl_2$ . *Journal of Synchrotron Radiation*, 17(2), 207-214. DOI: 10.1107/S0909049509048705.
- Flis, J., Manecki, M., & Bajda, T. (2011). Solubility of pyromorphite  $Pb_5(PO_4)_3Cl$  - mimetite  $Pb_5(AsO_4)_3Cl$  solid solution series. *Geochimica et Cosmochimica Acta*, (in press). DOI: 10.1016/j.gca.2011.01.021.
- Hettiarachchi, G. M., Pierzynski, G. M., & Ransom, M. D. (2000). In situ stabilization of soil lead using phosphorous and manganese oxide. *Environmental Science and Technology*, 34(21), 4614-4619. DOI: 10.1021/es001228p.
- Hettiarachchi, G.M., Pierzynski, G.M., & Ransom, M.D. (2001). In situ stabilization of soil lead using phosphorous. *Journal of Environmental Quality*, 30(4), 1214-1221. DOI: 10.2134/jeq2001.3041214x.
- Lenoble, V., Deluchat, V., Serpaud, B., & Bollinger, J.C. (2003). Arsenite oxidation and arsenate determination by the molybdene blue method. *Talanta*, 61(3), 267-276. DOI: 10.1016/S0039-9140(03)00274-1.
- Ma, Q.Y., Traina, S.J., & Logan, T.J. (1993). In situ lead immobilization by apatite. *Environmental Science and Technology*, 27(9), 1803-1810. DOI: 10.1021/es00046a007.
- Ma, Q.Y., Traina, S.J., Logan, T.J., & Ryan, J.A. (1994). Effects of aqueous Al, Cd, Cu, Fe(II), Ni, and Zn on Pb immobilization by hydroxyapatite. *Environmental Science and Technology*, 28(7), 1219-1228. DOI: 10.1021/es00056a007.
- Ma, L.Q., & Rao, G.N. (1997). Effects of phosphate rock sequential chemical extraction of lead in contaminated soils. *Journal of Environmental Quality*, 26(3), 788-794. DOI: 10.2134/jeq1997.00472425002600030028x.
- Manecki, M., Bogucka, A., Bajda, T., & Borkiewicz, O. (2006). Decrease of Pb bioavailability in soils by addition of phosphate ions. *Environmental Chemistry Letters*, 3(4), 178-181. DOI: 10.1007/s10311-005-0030-1.
- Manecki, M., & Maurice, P.A. (2008). Siderophore promoted dissolution of pyromorphite. *Soil Science*, 173(12), 821-830. DOI: 10.1097/SS.0b013e31818e8968.
- Manecki, M., Maurice, P.A., & Traina, S.J. (2000). Kinetics of aqueous Pb reaction with apatites. *Soil Science*, 165(12), 920-933.
- Melamed, R., Cao, X., Chen, M., & Ma, L.Q. (2003). Field assessment of lead immobilization in a contaminated soil after phosphate application. *Science of the Total Environment*, 305(1-3), 117-127. DOI: 10.1016/S0048-9697(02)00469-2.
- Miretzky, P., & Fernandez-Cirelli, A. (2008). Phosphates for Pb immobilization in soils: a review. *Environmental Chemistry Letters*, 6(3), 121-133. DOI: 10.1007/s10311-007-0133-y.
- Moore, D. M., & Reynolds Jr, R. C. (1997). *X-Ray Diffraction and the Identification and Analysis of Clay Minerals*. New York: Oxford University Press.
- Mozgawa, W., Król, M., & Bajda, T. (2009). Application of IR spectra in the studies of heavy metal cations immobilization on natural sorbents. *Journal of Molecular Structure*, 924-926, 427-433. DOI: 10.1016/j.molstruc.2008.12.028.
- Raicevic, S., Kaludjerovic-Radoicic, T., & Zouboulis, A.I. (2005). In situ stabilization of toxic metals in polluted soils using phosphates: theoretical prediction and experimental verification. *Journal of Hazardous Materials*, 117(1), 41-53. DOI: 10.1016/j.jhazmat.2004.07.024.
- Ranatunga, T.D., Taylor, R.W., Schulthess, C.P., Ranatunga, D.R.A., Bleam, W.F., & Zenwo, Z.N. (2008). Lead sorption on phosphate-pretreated kaolinite: Modeling, aqueous speciation, and thermodynamics. *Soil Science*, 173(5), 321-331. DOI: 10.1097/SS.0b013e31816d1e25.
- Ruby, M.V., Davis, A., & Nicholson, A. (1994). In-situ formation of lead phosphates in soils as a method to immobilize lead. *Environmental Science and Technology*, 28(4), 646-654. DOI: 10.1021/es00053a018.



- Ryan, J.A., Zhang, P., Hesterberg, D., Chou, J., & Sayers, D.E. (2001). Formation of chloropyromorphite in a lead-contaminated soil amended with hydroxyapatite. *Environmental Science and Technology*, 35(18), 3798-3803. DOI: 10.1021/es010634l.
- Sauvé, S., Martinez, C.E., McBride, M.B., & Hendershot, W.H. (2000). Adsorption of free lead ( $Pb^{2+}$ ) by pedogenic oxides, ferrihydrite, and leaf compost. *Soil Science Society of America Journal*, 64(2), 595-599. DOI: 10.2136/sssaj2000.642595x.
- Scheckel, K.G., & Ryan, J.A. (2002). Effects of aging and pH on dissolution kinetics and stability of chloropyromorphite. *Environmental Science and Technology*, 36(10), 2198-2204. DOI: 10.1021/es015803g.
- Stack, A.G., Erni, R., Browning, N.D., Casey, W.H. (2004). Pyromorphite growth on lead-sulfide surfaces. *Environmental Science and Technology*, 38(21), 5529-5534. DOI: 10.1021/es049487s.
- Taylor, R.W., Bleam, W.F., Ranatunga, T.D., Schulthess, C.P., Senwo, Z.N., & Ranatunga, D.R.A. (2009). X-ray absorption near edge structure study of lead sorption on phosphate-treated kaolinite. *Environmental Science and Technology*, 43(3), 711-717. DOI: 10.1021/es8020183.
- Xie, L., & Giammar, D.E. (2007). Equilibrium solubility and dissolution rate of the lead phosphate chloropyromorphite. *Environmental Science and Technology*, 41(23), 8050-8055. DOI: 10.1021/es071517e.
- Zhang, P., & Ryan, J.A. (1998). Formation of pyromorphite in anglesite-hydroxyapatite suspensions under varying pH conditions. *Environmental Science and Technology*, 32(21), 3318-3324. DOI: 10.1021/es980232m.
- Zhang, P., & Ryan, J.A. (1999a). Formation of chloropyromorphite from galena (PbS) in the presence of hydroxyapatite. *Environmental Science and Technology*, 33(4), 618-624. DOI: 10.1021/es980314a.
- Zhang, P., & Ryan, J.A. (1999b). Transformation of Pb(II) from cerrusite to chloropyromorphite in the presence of hydroxyapatite. *Environmental Science and Technology*, 33(4), 625-630. DOI: 10.1021/es980268e.
- Zhang, P., Ryan, J.A., & Bryndzia, L.T. (1997). Pyromorphite formation from goethite adsorbed lead. *Environmental Science and Technology*, 31(9), 2673-2678. DOI: 10.1021/es970087x.