

Anna CHRZAN^{1*} and Maria MARKO-WORŁOWSKA¹

PINE BARK AS INDICATOR OF SELECTED ANTHROPOGENIC POLLUTANTS

KORA SOSNY WSKAŹNIKIEM WYBRANYCH ZANIECZYSZCZEŃ ANTROPOGENICZNYCH

Abstract: Environmental monitoring of potentially toxic trace elements is important to control their concentrations in the environment. The suitability of bark and topsoil for monitoring of these heavy metals and acidifying gases pollution was investigated. Concentrations of Pb, Cd, Ni were determined in topsoil and in samples of necrotic bark of *Pinus sylvestris* L. collected along transects around the Skawina industry center and in parts of Bielansko-Tyniecki Landscape Park in Krakow.

After comparing the concentration of heavy metals and pH value in pine bark with topsoil, it was observed that topsoil is better biomonitor for lead and nickel than bark of *Pinus sylvestris* and that bark appear to be suitable bioindicator of atmospheric deposition only for cadmium and acidifying components. The results obtained confirm the negative impact of aluminium work, power plant and transport on quality and number of environmental pollutants at the sites situated near these industries and road with intensive traffic. Therefore, the constant monitoring of these localities is necessary.

Keywords: *Pinus sylvestris* L., necrotic bark, heavy metals, topsoil, biomonitoring

Introduction

Heavy metals (also called potentially trace elements PTEs) are very important parts of the Earth's crust. The environmental problem of which they are a cause is complex. On the one hand, a large decrease of russets of which they are integral parts can be noticed, whereas on the other hand, they occur in higher and higher concentrations in the living organisms and in the environment [1]. In terrestrial ecosystems topsoil is the main source of heavy metals for the organisms. At the same time, soil has filtering and buffering functions, facilitates decomposition of toxic substances and protects the environment against the extensive transmission of trace elements. These soil qualities

¹ Department of Ecology and Environmental Protection, Institute of Biology, Pedagogical University of Krakow, Podbrzezie 3, 31-054 Kraków, Poland, phone: +48 12 662 66 81, fax: +48 12 662 66 82, email: annachrn871@gmail.com

* Corresponding author: annachrn871@gmail.com

are persistent as long as the balance in their biological functioning is maintained. In case of dry deposition of dusts – that are the carriers of metals in the environment-metals remain practically inactive and are absorbed mechanically. Whereas in case of the presence of high concentrations of acidifying gases and simultaneous rainfalls, the growth of bioaccumulation of more easily soluble ionic forms of metals occurs in the environment [2].

Cationic trace elements are maintained in surface soils and their extensive accumulation is the primary cause of chemical degradation of soils. The soil polluted by heavy metals can transmit them to subsequent links of the food chain – to plants, animals and man, or it can be the source of secondary contamination of air and waters and as such, can influence man, disregarding the food chain [3].

Intensive development of industry (metallurgy, burning of energetic raw materials) and steadily increasing car transport constitute the main source of contamination by heavy metals and acidic compounds (SO_2 , SO_3 and NO_x) of soils adjacent to the industrial areas and traffic paved roads.

The consequence of the accumulation of metals and their toxic impact is biological deactivation of the soil subsystem. Because of the decrease in density, biodiversity and activity of the microorganisms and of pedofauna, limitations of the processes of organic substances decomposition occur. Additionally, terrestrial ecosystems productivity decreases [3, 4]. Heavy metals and acidifying compounds are one of the most durable and toxic contaminates of the soil subsystem.

The anthropogenic processes that cause the emission of acidifying compounds are the cause of the spreading of many metals and the growth of their toxic impact on living organisms [5]. Respectively, the content of heavy metals (especially Pb, Ni or Cd) in soils can be treated as an indicator of the environmental degradation of different branches of the industry and car transport.

Tree bark is an ideal natural absorbent as it is dead tissue that does not grow anymore. The pine bark surface is very porous and the absence of metabolic processes makes it highly unreactive for inorganic and organic substances [6, 7]. The changes in the chemical composition of the surface layers can be documented. Kuik and Wolterbeek [8] and another [9, 10], proposed the use of tree bark samples as biomonitoring of heavy metals. But also the pine bark is a good indicator of air pollution with cement-lime dust [11, 12].

Exposed to the air for years, it is an indicator for the air pollutant concentration. By analyzing the pollutant concentrations in bark, the environmental air pollution surrounding the trees can be determined. Monitoring of the environment with the use of bark not only enables evaluation of present layer of the accumulation of elements and polluting compounds, but also evaluation of the course of the process in the previous years. Adjacent trees are supposed to demonstrate a similar degree of pollution accumulation and may be used consequently to determine the extent of the pollutants. Similarly the results on the bark of model trees analysis taken on surfaces at different distances from source emission can illustrate the problem of the spread of pollutions [12]. Tree bark is a good bioindicator because it remains in place for an extended period of time, it is easily accessible and sampling does not damage the tree [10]. What is

more, sampling and analysis of necrotic bark is fast and economical. Kuik and Wolterbeek [8] study demonstrated the potential of bark as a biomonitor on a larger scale.

In this investigation, the hypotheses to be tested were:

– atmospheric deposition is the main cause for the presence of heavy metals and acidifying compounds in topsoil and in the bark of *Pinus sylvestris* investigated in the Skawina industrial area and in the parts of Bielansko-Tyniecki Landscape Park in Krakow located in southeast Poland,

– topsoil and pine bark are suitable for biomonitoring pollution by the industry and car transport.

The aim of the present work was the recognition of contamination by Pb, Cd and Ni and acidifying compounds of the surface soil (topsoil) and tree bark of approximately similar aged trees at localities situated at different distances from industrial and transport emitters.

The results of these researches can provide important information on condition of the environment as well as of environmental processes.

Materials and methods

A. Characteristic of area analyzed

Sites of samples collection was chosen with the regard to:

– lawns in ecotonic area of forests: (topsoil samples), and around 30 year-old *Pinus sylvestris* L. (necrotic bark),
– direction of predominant winds,
– localizations of particular emitters.

For the research 5 sites (1–5), were chosen in compliance with the research objectives proposed. The samples were taken in spring 2012 at 5 different distances from the source of emission. The 3 sites (1, 2, 3) examined are situated in Skawina – a city characterized by heavy industrialization (aluminium-works, power plant and others), densely populated and situated approximately 18 km southwest from the center of Krakow and the 2 sites (4, 5) situated in the Bielansko-Tyniecki Landscape Park of Krakow – park encompassed by the Krakow city limits.

Site 1 – located closely to the industrial works (300 m south west from the power plant and 700 m north east from the aluminium works and situated about 5 meters north of a heavy used road).

Site 2 – situated around 2600 m and around 2400 m south east from the power plant and aluminium works,

Site 3 – situated 1200 m south west from the power plant and 450 m north west from the aluminium works and approximately 200 m from the road.

Site 4 – situated 4 km north east from the power plant and 5 km from the aluminium works.

Site 5 – situated 5 km north east from the power plant and 6 km from the aluminium works, and around 200 m west from the ring road.

The sites of the samples taken are marked in Fig. 1.

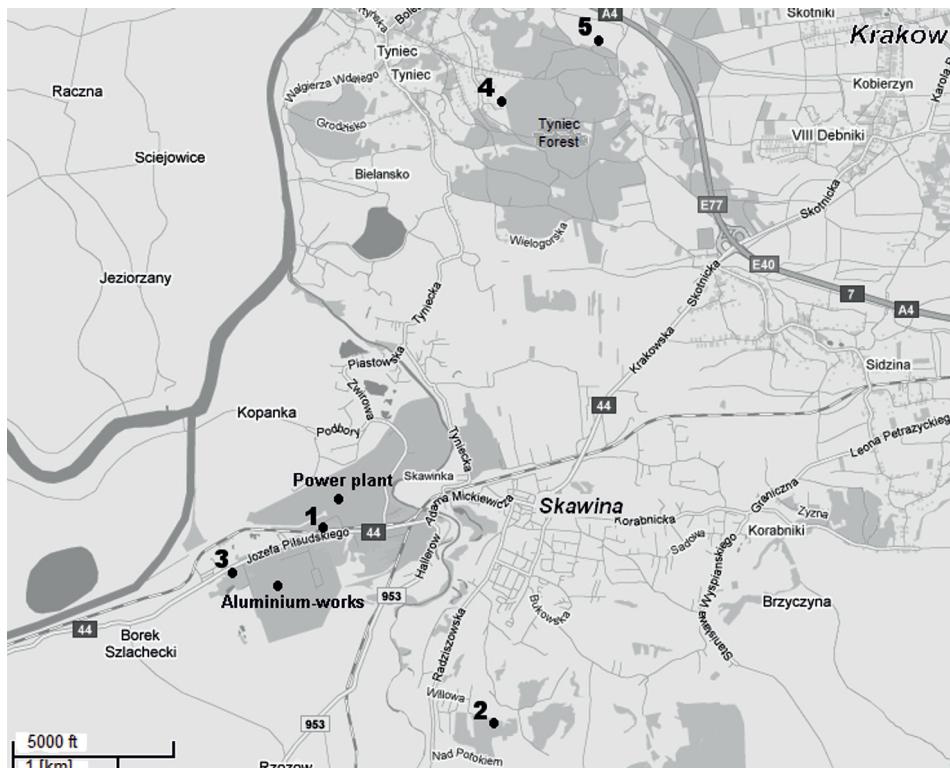


Fig. 1. Sites of samples taking

At each sites four representative topsoil samples were taken at ecotonic zones of the forests.

B. Samples collection and analysis

For evaluation of pollutions of air and soil by heavy metals and acidifying compounds necrotic pine bark (*Pinus sylvestris* L.) and topsoil from the pines surroundings were used as the accumulative cumulative bioindicators. The temperature, soil moisture and pH of topsoil and bark were examined and the content of Pb, Cd and Ni was determined. The topsoil reaction (pH) was examined on place with a pH-meter-WTW pH 330, placed into the soil to a depth of 5 cm. The topsoil samples were taken by forcing a 5 cm diameter steel cylinder to a depth of 5 cm. At each site the 4 soil samples were taken at eastern side (before the trees) and the 40 samples of bark 3 mm thick from the trunk of 4 trees at the high of around 1.5 m from the soil, 5 in the eastern side of each tree and 5 on the western side. The trees analyzed were from 2 m to

4 m distance from each other, had similar breast height diameter (35 cm to 40 cm) and had a similar age (around 30 years).

The samples were dried in a temperature of 65 °C for 3 hours, then milled to a powder. One gram of powdered bark from each sample was weighed and mixed with 5 cm³ of distilled water. After 48 hours the pH reaction was measured using a ELMETRON type CPC-401 pH-meter. Chemical analyses of heavy metals in the topsoil and bark were conducted by mapping the content of the general forms of Cd, Pb, Ni, determined by the FAAS method. Dried samples of topsoil and bark (1 g each) were mineralized. For this purpose dried samples of soil and pine bark were mixed with 3 cm³ of 65 % HNO₃, the solution was heated to a temperature of 110 °C and left for 4 hours to cool. The filtered liquid was poured into measuring flasks and filled with distilled water. In the solutions of soil and bark prepared, the content of heavy metals was determined by atomic absorption spectrometer (AAS-Cole-Parmer, BUCK 200A).

Statistical analysis

As variances showed no parametrical distribution (Shapiro-Wilk test, $p > 0.05$), we used non parametric Kruskal-Wallis ANOVA [13]. The correlations between metal concentrations, pH, and sites were calculated using the Pearson correlation coefficient and Student's test. Statistical significance was defined at $p < 0.05$. All analyses were performed using STATISTICA 12 computer program.

Results and discussions

The pH of water extracts of pine bark and pH of topsoil

The pH value is subjected to the quickest changes when exposed to external factors and, among others, it directly and indirectly influences many mechanisms releasing heavy metals in soils [14]. In normal conditions the pH of *Pinus sylvestris* bark is 3.5 [9], whereas the average pH of pine bark at sites analyzed indicated high acidity and fluctuated from 2.68 to 3.44 (Table 1).

Comparison of the pH reaction indicated considerably higher acidity of pine bark versus topsoil at each sites, and these differences were statistically significant (Table 2).

The average value of soil pH fluctuated from 4.31 to 6.04 (Table 1). The highest acidity (pH lower than 4.5) was observed in soils at sites 2 and 3). At the same time similar, almost parallel increasing or decreasing tendencies of pH of bark and topsoil were noticed regarding all sites. The pH value of pine bark on the eastern and on the western side did not differ much, however at sites 1, 2, 3 situated at the closest distance and west of industrial emitters, a slightly higher acidity on the eastern side of trunk was observed (Table 1).

The lowest average pH value of bark and of soil was recorded at site 2 (Table 1). It is probably that here polluted air touched with the north-east winds from the north-east part of Krakow, where is located a steelworks and electro-thermal power plant. Site 5 situated furthest away from the industrial emitters, but approximately 200 m from a ring

road was characterized by high bark acidity, as well as the highest content of Pb in the topsoil (Fig. 2). This result proves a big impact of transport pollutions flowing with frequently occurred natural wind currents blowing from the ring road.

Table 1

Acidity and its range of changes (in bracket) of pine bark and topsoil in relation to cardinal directions

Parameter	Cardinal directions	Locality 1	Locality 2	Locality 3	Locality 4	Locality 5
Pine bark pH	East	3.14 (2.98–3.24)	2.68 (2.60–2.82)	3.18 (2.63–3.94)	3.29 (3.05–3.46)	3.12 (3–3.23)
	West	3.28 (3.05–3.59)	2.99 (2.61–3.49)	3.44 (2.78–3.78)	3.14 (3.01–3.27)	3.02 (2.88–3.23)
Topsoil pH	East	6.04 (5.34–6.79)	4.25 (3.79–5.47)	4.4 (4.01–5.03)	4.61 (4.37–5.05)	4.69 (4.52–5.21)

Table 2

The statistical significance of *p*-levels related to comparison of pH and metal's concentration in soil and pine bark

Parameters	pH	Pb	Cd	Ni
Pine bark East-West	**	**	*	**
Pine bark East-Topsoil	**	**	**	**
Pine bark West-Topsoil	**	**	ns	**

* 0.05 > p > 0.01; ** 0.01 > p > 0.001; ns – not significant.

The low soil pH is negative from the ecological perspective – in acidic soil the mobility of heavy metals (mainly Cd and Ni) increases which facilitates transfer of these metals into the circulation of matter. The results of presented research indicate that the pH value of pine tree bark must be regarded as an indicator for air acidification.

The content of potentially toxic trace elements (PTEs)

Amongst the metals examined Ni is the only one that is indispensable (for humans 25–35 µg/day) for the proper functioning of the organisms. Its overdose, however, can have very negative consequences. Cd and Pb do not have any positive influences on organisms and constitute some of the most dangerous heavy metals; the accepted dose determined by the WHO is: Pb – 415–550 µg/day, Cd – 57–71 µg/day [1]. The occurrence of heavy metals in soil and plants can be treated as an indicator of the quality of the environment and *Pinus silvestris* is commonly used in fitoindication [15]. Intensity and scope of the contamination of soils and plants by trace elements depend on many factors at local conditions. Additionally, the important source of heavy metals accumulated on the surface of soils and plants is so-called long distance emission related to the spreading of pollutions over large distances.

The climatic conditions (rainfalls, temperature, direction and wind speed) play an important role in dispersion and deposition of air pollutants. For an estimation of the environmental health and threat of heavy metals, an evaluation of their mobility, that is, possibility of their transmission to biogeochemical circulation, is necessary [16]. High mobility of Cd, Ni and the relatively low mobility of Pb in soils is regulated by their pH and modified by the other soil qualities such as content of organic substances, occurrence of other metals and humidity. In oxidative environments with acidic reactions, Cd and Ni are characterized by high mobility, to the lesser extent the mobility of Pb increases as well [17].

Among metals measured in topsoil the concentration of Pb was the highest and its amount was higher than the average natural occurrence in the soils of Poland [18] (Fig. 2).

Lead – a component of commonly used, until recently, leaded petrol is the contamination attributed to transport and to various industrial activities, for example metallurgy. Pb was used in leaded petrol in amounts ranging from 0.15 to 0.30 g/dm³ as an antidetonant increasing the anti-knock index. Pb occurs in unleaded petrol as well, where its content cannot exceed the amount of 0.005 g/dm³. The other main sources of Pb emission are coal burning and metallurgical industries. It is a metal typical for low emitters and non-organized emissions. Pollutions from this source do not spread over long distances [19]. According to the literature, [1], natural concentrations of this metal should not exceed 20 mg/kg for most soils.

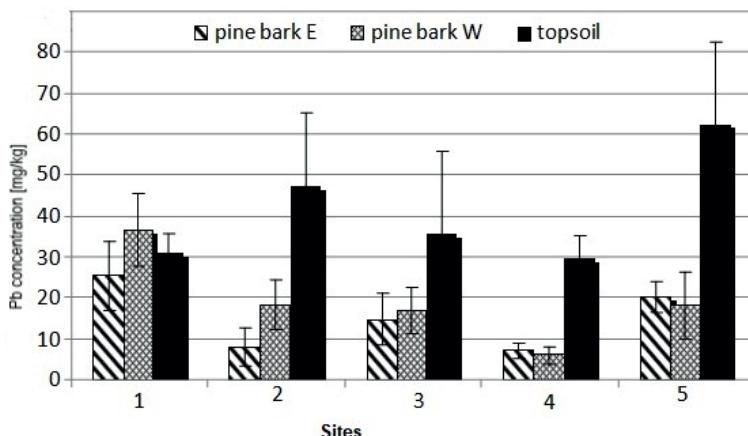


Fig. 2. Pb concentration

The highest Pb content was observed in topsoil at site 5 situated only approximately 200 m away from the ring road and in bark at site 1 adjacent to the busy road (Fig. 2). The Pb content in soils analyzed fluctuated from 29.37 mg/kg at site 4 furthest away from industrial emitters, ring road and busy road to 62.26 mg/kg at site 5 which is remote from industry, but situated close to a ring road. The concentrations of Pb in topsoil and bark of site 4 were significantly lower than those of sites 1, 2, 3, 5 (Fig. 2).

Soils of habitats analyzed contain slightly higher amount of Pb than the natural average content in soils in Poland. However, only in soil at site 5 the amount of Pb is averagely 2 times higher than in soils of remaining sites and exceeds acceptable content according to the standards of soils [18] and earths qualities. The greatest effect of traffic Pb pollution on the soil subsystem was observed up to a distance of 200 m away from the motorway (Fig. 2) which proves higher influence of busy roads than in the results presented by Ruhling and Tyler [20]. Their studies of Pb accumulation in many plants growing next to busy roads in Sweden, show the most intensive accumulation of Pb up to 150m from the road. Observations of Jaworska and Murowana [21] stating the impact of distance up to 300 m from the road on plants are similar to the results of presented research.

Concentration of Pb general forms in topsoil compared with Pb concentration in pine bark is much higher at sites 2, 3, 4 and 5 (Fig. 2). The content of Pb spanning from 2 to more than 6 times higher in soil than in bark was detected here. The exception is site 1, situated closest to the industrial emitters and busy road, that is characterized by slightly higher content of Pb in the bark on the western side of the trunk, that is on the side of the aluminum work (Fig. 2). Simultaneously, at this locality the highest Pb content in the bark can be noticed. The highest Pb concentration in pine bark at site 1 can be explained by the proximity to the heavy traffic road. With increasing distance from the emission center (car transport and power plant) the Pb content of bark decreases.

The lowest Pb content in bark as well as in topsoil, was observed at site 4 situated furthest away from roads (Fig. 2). Therefore it is probable that the main emitter of Pb on area analyzed is car transport. In cases sites 1, 2, 3 the average concentration of lead on the eastern side pine bark is lower than on the western side, and they differ significantly (Table 2).

Cadmium, that is emitted by metal works, is one of the most dangerous heavy metals for the environment [22]. What is more, along roads, the source of contaminations are for example greases used in vehicles, tyre abrasion and other parts of vehicles. Cd is accumulated mainly in surface soil levels and the higher the reactivity (pH) of soil, the more readily it is absorbed by it. Cadmium is more easily activated and mobile in soils of pH 4.5–5.5. With higher pH values Cd is immobilized forming carbonates [23]. In soils examined it occurs in natural amounts or slightly elevated ones and fluctuated from 0.41 to 0.97 mg/kg (Fig. 3). In 6-degree IUNG classification [18] 0.5 mg/kg d.m. is considered as natural amount of Cd in soil. On this basis it was stated that topsoils at sites 1, 4 and 5 exceed this amount (Fig. 3).

The highest Cd concentration was detected in soil with the least situated near a ring road. Whereas in tree bark the highest Cd content was detected at site situated nearest to the aluminium works and power plant (Fig. 3), and the highest content of the general forms of this element was noted on the eastern side of tree-trunks. Higher Cd content on the eastern side of tree-trunk (differently than in case of Pb) can prove that the main source of emission of this element is industry (Fig. 2, 3). As the research indicate, Cd accumulates best in bark – according to Samecka-Cymerman et al. [24] they state that bark is a better accumulator of Cd compared to *Pinus schreberi*, whereas in soil its

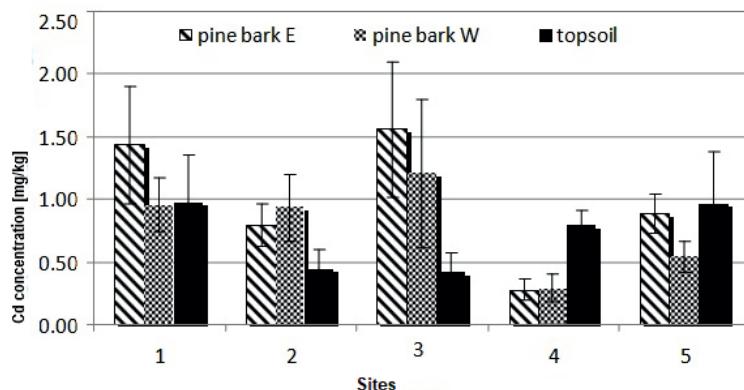


Fig. 3. Cd concentration

content is considerably lower. However in soils at sites situated near roads (1 and 5) the Cd content is similar to the content observed in barks at the same localities.

In spite of the statement presented in literature that emissions released at the height [19] can be observed in case of Cd and Ni, because they are emitted mainly from tall smokestacks. Some seasonal changes in concentrations of metals accumulated in moss and lichens were also indicated [25]. There was observed, i.a., an increase in Cd concentration at the beginning of the growing season, which may be related to low emissions during the heating season. The results prove that apart from industrial emissions, car transport influences heavily the content of Cd. The results obtained by Marko-Worlowska et al. [26] confirm at the same time, that motorways are the main emitters of Pb and big emitter of Cd, in the immediate vicinity nearest and up to 200 m away from the motorway.

Nickel is emitted mainly by the metallurgical industry and during coal burning and petrol use [27]. In the case of Ni considerably higher amounts were observed in soils at

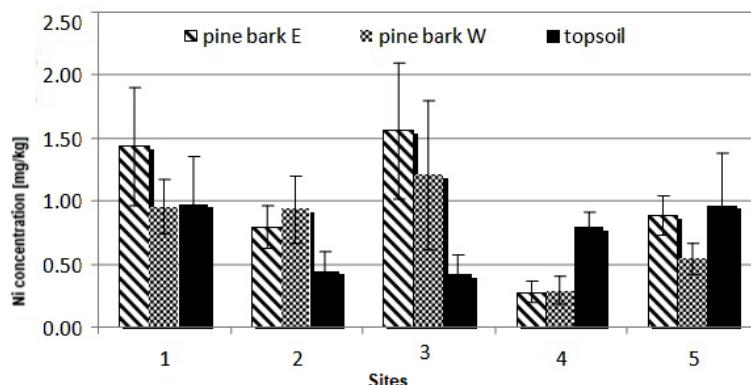


Fig. 4. Ni concentration

all sites analyzed (from 2.73 to 4.41 mg/kg), whereas in barks the amount was from 0.4 to 1.9 mg/kg (Fig. 4). Natural concentrations of this metal in soil are 25 mgNi/kg d.m. [14]. In all soils analyzed its content was much lower. In tree-bark the highest content of Ni (different than in case of Pb, and similar to the case of Cd) was noted on the eastern side of the trees (Fig. 4). It proves the significant influence of industry on Ni emissions.

Results of statistical analysis comparing accumulated metals concentrations and pH value in bark of *Pinus sylvestris* in relation to topsoil concentration and distance from the source of pollution are presented in Tables 2–4. Important correlations between Pb and Cd concentration was stated determined in bark at sites 1–4 and in topsoil at sites 2 and 5. At site 5 correlation between Pb and Ni as well as between Cd and Ni was observed. In most circumstances, a positive correlation is observed between lead and cadmium concentrations in soil and pine bark, which predicts the higher the lead content the higher the cadmium concentration (Table 3).

Table 3

Correlation between pH (topsoil and bark) and heavy metals concentration on each sites

Parameters	<i>r</i>				
	Site 1	Site 2	Site 3	Site 4	Site 5
Pine bark East Cd-Pb	0.86**	0.72**	0.68**	-0.66**	
Pine bark East Ni-Pb			-0.89**		0.55*
Pine bark East Cd-pH			-0.61*	0.63**	
Pine bark East Pb-pH					
Pine bark West Ni-Pb					-0.54*
Pine bark West pH-Pb	-0.51*	0.56*	-0.76**	-0.52*	-0.55*
Pine bark West pH-Cd		0.60*			-0.62*
Topsoil Pb-Cd		0.89**			0.57*
Topsoil Cd-Ni					0.71**
Topsoil Pb-Ni					0.82**
Topsoil pH-Ni		0.89**	0.58*		
Topsoil pH-Cd			0.51*		

* 0.05 > *p* > 0.01; ** 0.01 > *p* > 0.001.

Concentrations of Cd and Pb in topsoil and in bark were influenced by the distance from industrial emitters and busy roads. Concentration of Ni in topsoil was not influenced by these distances, whereas concentrations of Cd, Pb, Ni in bark were not influenced by concentration of this element in soil. Important influence of factors: direction of the wind and site and interactions between pH value and heavy metals concentration was stated (Tables 3, 4).

Table 4

Results of ANOVA comparing – multiple factor analysis (p -value)

Factor	pH	Pb	Cd	Ni
Direction	ns	***	*	***
Sites	***	***	***	***
Interaction (Direction and Locality)	*	***	*	***
Locality	***	***	***	***
Material	***	***	**	***
Interaction (Locality and Material)	***	***	***	***

* $0.05 > p \geq 0.001$; ** $0.001 > p \geq 0.0001$; *** $p < 0.0001$; ns – not significant.

Conclusions

The results obtained confirm the negative impact of aluminium work, power plant and transport on quality and number of environmental pollutants at the sites situated near these industries and road with intensive traffic. Therefore, the constant monitoring of these localities is necessary.

Significant differences in pH values and concentrations of lead, cadmium, nickel in the necrotic bark of pine trees and topsoil at 5 sites were observed. In addition, the spread of contamination is effected by the direction of the wind that moves the pollution long distances.

The results show that a better absorber of lead and nickel is topsoil, whereas pine bark accumulates more cadmium and pollutions caused by acidifying gases:

- a) topsoil is better biomonitor for lead and nickel than pine bark,
- b) pine bark appears to be better bioindicator of atmospheric acidifying compounds and cadmium deposition and with comparison with another bioaccumulators can be suitable biomonitor for environmental pollution by heavy metals.

References

- [1] Kabata-Pendias A, Pendias H. Trace Elements in Soils and Plants. 3rd edn. CRC Press; 2001. ISBN 0-8493-1575-1
- [2] Morselli L, Olivieri P, Brusori B, Passarini F. Soluble and insoluble fractions of heavy metals in wet and dry atmospheric depositions in Bologna, Italy. Environ Pollut 2003;124:457-469. DOI:10.1016/S0269-7491(03)00013-7.
- [3] Kabata-Pendias A. Soil-plant transfer of trace elements – an environmental issue. Geoderma. 2004; 122(2-4): 143-149. <https://doi.org/10.1016/j.geoderma.2004.01.004>
- [4] Howard JL, Shu J. Sequential extraction analysis of heavy metals using a chelating agent (NTA) to counteract resorption. Environ Pollut. 1996;91:89-96. DOI: 10.1016/0269-7491(95)00023-K.
- [5] Niesiobędzka K. Mobile forms and migration ability of Cu, Pb and Zn in forestry system in Poland. Environ Earth Sci. 2016;75:1-8. DOI: 10.1007/s12665-015-4821-9.
- [6] Schulz H, Popp P, Huhn G, Stärk H-J, Schürmann G. Biomonitoring of airborne inorganic and organic pollutants by means of pine tree barks. – I. Temporal and spatial variations. Sci Total Environ. 1999;232;1:49-58. DOI: 10.1016/S0048-9697(99)00109-6.

- [7] Sawidis T, Breuste J, Mitrovic M, Pavlovic P, Tsigardas K. Trees as bioindicator of heavy metal pollution in three European cities. *Environ Pollut.* 2011;159:3560-3570. DOI: 10.1016/j.envpol.2011.08.008.
- [8] Kuik P, Wolterbeek HT. Factor-analysis of trace-element data from tree-bark samples in the Netherlands. *Environ Monit Assess.* 1994; 32: 207-226 DOI: 10.1007/BF00546277.
- [9] Grodzińska K. Tree bark – sensitive biotest for environment acidification. *Environ Intern.* 1979;2(3):173-176. DOI: 10.1016/0160-4120(70)90075-8.
- [10] Harju L, Saarela KE, Rajander J, Lill JO, Lindroos A, Heselius SJ. Environmental monitoring of trace elements in bark of *Scots pine* by thick-target PIXE. *Nucl Instrum Methods B.* 2002;189:163-167. DOI:10.1016/S0168-583X(01)01031-X.
- [11] Sporek M. Seasonal variability of chemical composition of precipitation in the coniferous stands. *Proc ECOpole.* 2018; 12(2):571-577. DOI: 10.2429/proc.2018.12(2)059.
- [12] Sporek M. The use of bark of scots pine in environmental impact assessment of cement and lime dust. *Proc ECOpole.* 2016;10(2):505-509. DOI: 10.2429/proc.2016.10(1)054.
- [13] Sokal RR, Rohlf FJ. *Biometry: The principles and practices of statistics in biological research.* New York: WH Freeman and Company; 2012. ISBN: 0-7167-8604-4 or 978-0-7167-8604-7.
- [14] Filipek T, Skowronska M. Current dominant causes and effects of acidification of soils under agricultural use in Poland. *Acta Agrophys.* 2013;20(2):283-294. <http://produkcia.ipan.lublin.pl/uploads/publishing/files/Filipek-283-294.pdf>.
- [15] Kosiorek M, Modrzewska B, Wyszkowski M. Levels of selected trace elements in Scots pine (*Pinus sylvestris* L.), silver birch (*Betula pendula* L.), and Norway maple (*Acer platanoides* L.) in an urbanized environment. *Environ Monit Assess.* 2016;188(10):598. DOI: 10.1007/s10661-016-5600-0.
- [16] Losfeld G, L'Huillier L, Fogliani B, Mc Coy S, Grison C, Jaffré T. Leaf-age and soil-plant relationships: key factors for reporting trace-elements hyperaccumulation by plants and design applications. *Environ Sci Pollut Res Int.* 2015;22(8):5620-32. DOI:10.1007/s11356-014-3445-z.
- [17] Aydinald C, Marinova S. Distribution and forms of heavy metals in some agricultural soils. *Pol J Environ Stud.* 2003;12(5):629-633. <http://www.pjoes.com/pdf/12.5/629-633.pdf>.
- [18] Regulation of the Minister of the Environment on 9 September 2002 on the standards of the soil quality and ground quality. Limit values of concentration in the soil or ground. 1.09.2002. DzU. Nr 165, 1359. <http://isap.sejm.gov.pl/>
- [19] Liang J, Mao J. Source analysis of global anthropogenic lead emissions: their quantities and species. *Environ Sci Pollut Res Int.* 2015;22(9):7129-38. DOI: 10.1007/s11356-014-3878-4.
- [20] Ruhling A, Tyler G. Changes in the atmospheric deposition of minor and rare elements between 1975 and 2000 in south Sweden, as measured by moss analysis. *Environ Pollut.* 2004;131:417-423. DOI:10.1016/j.envpol.2004.03.005.
- [21] Jaworska M, Murowana D. Influence of environment pollution on entomofauna of city gardens. *Ecol Chem Eng A.* 2008;15(1-2):71-78.
- [22] Martiniakova M, Omelka R, Jancova A, Formicki G, Stawarz R, Bauerova M. Accumulation of risk elements in kidney, liver, testis, uterus and bone of free-living wild rodents from a polluted area in Slovakia. *J Environ Sci Health A.* 2012;47(9):1202-1206. DOI: 10.1080/10934529.2012.672062.
- [23] Alloway BJ. Heavy metals in soils: trace metals and metalloids in soils and their bioavailability. Dordrecht: Springer; 2010. ISBN 978-94-007-4469-1.
- [24] Samecka-Cymerman A, Kosior G, Kempers AJ. Comparison of the moss *Pleurozium schreberi* with needles and bark of *Pinus sylvestris* as biomonitor of pollution by industry in Stalowa Wola (southeast Poland). *Ecotoxicol Environ Saf.* 2006;65:108-117. DOI: 10.1016/j.ecoenv.2005.05.009.
- [25] Kłos A, Ziembik Z, Rajfur M, Dołęga-Ziembik A, Bochenek Z, Bjerke JW, et al. Using moss and lichens in biomonitoring of heavy-metal contamination of forest areas in southern and north-eastern Poland. *Sci Total Environ.* 2018;627:438-449. DOI: 10.1016/j.scitotenv.2018.01.211.
- [26] Marko-Worłowska M, Chrzan A, Łaciak T. Scots pine bark, topsoil and pedofauna as indicators of pollutions in terrestrial's ecosystems. *J Environ Sci Health A.* 2011;46(2):138-148. DOI: 10.1080/10934529.2010.500896.
- [27] Tian HZ, Lu L, Cheng K, Hao JM, Zhao D, Wang Y, et al. Anthropogenic atmospheric nickel emissions and its distribution characteristics in China. *Sci Total Environ.* 2012;417:148-157. DOI: 10.1016/j.scitotenv.2011.11.069.

**KORA SOSNY WSKAŹNIKIEM
WYBRANYCH ZANIECZYSZCZEŃ ANTROPOGENICZNYCH**

Instytut Biologii
Uniwersytet Pedagogiczny im. Komisji Edukacji Narodowej, Kraków

Abstrakt: Monitorowanie stężenia potencjalnie toksycznych pierwiastków śladowych w środowisku jest bardzo ważne. W tym celu zbadano przydatność kory sosny i wierzchniej warstwy gleby do monitorowania metali ciężkich i związków zakwaszających w środowisku. Stężenia Pb, Cd, Ni określano w glebie oraz w próbkach kory *Pinus sylvestris* L. zebranych wokół centrum przemysłowego Skawiny oraz na terenie Bielańsko-Tynieckiego Parku Krajobrazowego w Krakowie. Badania porównawcze wykazały, że gleba jest lepszym wskaźnikiem zanieczyszczenia ołowiem i niklem niż kora sosny, natomiast kora wydaje się być odpowiednim bioindykatorem tylko dla kadmu i związków zakwaszających. Uzyskane wyniki potwierdzają negatywny wpływ huty aluminium, elektrowni i transportu na jakość i wielkość zanieczyszczeń środowiska w miejscowościach położonych w pobliżu tych gałęzi przemysłu i drogach o intensywnym ruchu. Dlatego konieczne jest stałe monitorowanie tych miejsc.

Słowa kluczowe: kora martwicowa sosny, metale ciężkie, gleba, biomonitoring, zanieczyszczenia antropogeniczne